

Circular Economy of Packaging and Relativity of Time in Packaging Life Cycle

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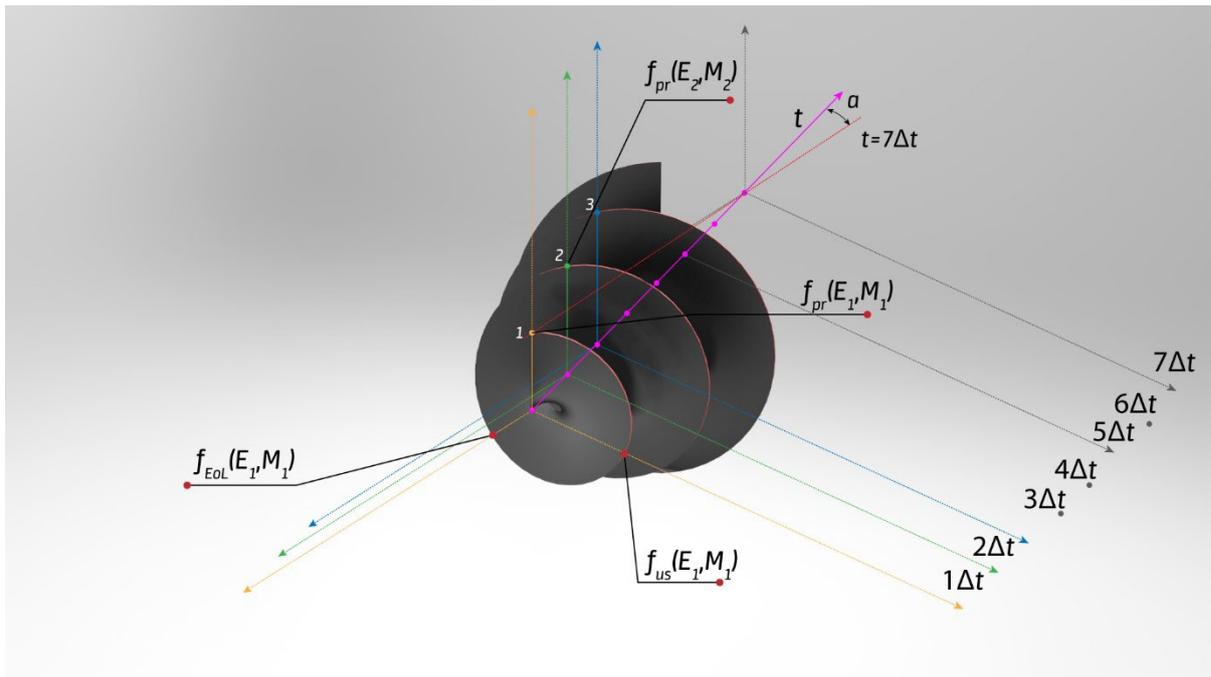
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33 GRAPHICAL ABSTRACT



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35 ABSTRACT

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37 The circular economy paradigm aims to improve the use of material and decrease the
38 negative impacts of the life cycle of products on the environment. In line with the broad variety of
39 conceptualizations and definitions used to describe the circular economy, there are also numerous
40 circularity indicators available in the literature. For analyses that focus on the speed of recycling, we
41 introduce a new methodology. This paper applies the three variables that define the value creation
42 principles in the widely accepted definition of circularity provided by the Ellen MacArthur Foundation:
43 material, energy and time. We show that including time in the LCA methodology may improve the
44 understanding of the system under study, especially for products such as packaging that have a
45 relatively short usage time in the technosphere compared to their recycling time. For this purpose we
46 develop a formula that includes the time necessary for obtaining the secondary material needed for
47 the production of "n + 1" product. The paper shows that we need to consider the production of
48 additional packaging products and that the quantity of these products depends on the time needed
49 for recycling in the waste management system. This aspect has traditionally been neglected when
50 developing comparative LCAs between systems that serve the same function. The proposed approach
51 to packaging LCA contributes to an important scientific debate over the allocation of credits and
52 burdens between several consecutive life cycles of a material.

53

54 Key words

55 Circularity Metrics, Time, LCA, Life Cycle

1. Introduction

Over the past 50 years, the European Union (EU) has placed particular emphasis on the need for efficient and sustainable usage of resources, beginning with the adoption of the first EU Environment Action Programme in 1972 (Hey, 2005). The first legislative change related to packaging waste was introduced in 1985 with the Council Directive on containers of liquids for human consumption (Council of the European Union, 1985), since when the EU has introduced several mandatory legislation changes for its member states related to packaging (European Commission, 2018a; European Union, 1994). Based on these legislative requirements, the percentage of recycled and recovered packaging material in the EU has been steadily increasing (Eurostat, 2018). However, the amount of packaging waste generated has been increasing at the same pace (Eurostat, 2018).

The introduction of the circular economy paradigm has changed the European Commission (EC)'s approach to and stimulated development of strategic actions for reducing packaging waste. The EC's first Circular Economy Action Plan (2015–2020) had the following three aims: (i) introducing circular design and circular production processes to packaging; (ii) empowering customers by focusing on product and organizational environmental footprint; and (iii) changing the legislative framework for waste and closing the loop on recovered packaging materials (European Commission, 2019a). With the adoption in 2019 of the so-called "Single-use Plastics Directive" and its circularity principles, the EC introduced a systematic sectoral approach to packaging for the first time (European Commission, 2019b). The new Circular Economy Action Plan adopted in 2020 aims to reduce (over)packaging and packaging waste by promoting new designs for re-use and recyclability, reducing the complexity of packaging materials, introducing labelling for correct separation, and revising the regulations on the shipment of waste (European Commission, 2020).

Life cycle assessments (LCA) are commonly conducted by industry and scientists as a comprehensive tool for assessing the environmental impacts of different packaging products (Batlle-Bayer et al., 2019; Ferrara and De Feo, 2018; Gomes et al., 2019; Molina-Besch et al., 2019). A specific challenge in monitoring the circularity of packaging relates to the application of appropriate circularity metrics. The literature divides these metrics into three groups according to the robustness of the analysed systems: micro-scale (product, company); mezzo-scale (industrial systems); and macro-scale (city, regional, national or beyond) (Geng et al., 2012; Janik, 2017).

The past decade has seen intense scientific debate regarding possible solutions for calculating the circularity of products and services, with a strong focus on the role of packaging. Among the important outcomes of this scholarship, García-Barragán et al. (2019) have developed a mathematical approach for comparing the linear economy and the circular economy and defining circular economic growth, while Di Maio et al. (2017) have proposed a market value approach that calculates the value-based resource efficiency of products and services that are returned after the end of life as a metric for circularity.

Numerous methods and indicators have been devised for assessing circularity, especially the circularity of products, on a micro-scale. For example, Pauliuk et al. (2017) have extended the regional scope of the MaTrace model to follow steel flows and assess losses at the end of life through multiple product life cycles. Specifically, they have developed a circularity index that takes account of the purity, quality and recoverability of the material present within the system within a fixed time interval. Linder et al. (2017) developed a circularity indicator and practically combined this with the environmental impacts, while a later study by Linder et al. (2020) developed an indicator calculated as the ratio between the economic value of the recirculated components and the economic value of the entire product, thus taking into consideration all the aspects of recirculation, i.e. reusability, remanufacturing, repurpose and recycling. In addition, a recent study by Cordella et al. (2020) has

103 assessed material efficiency as an aspect contributing to the circular economy in the existing Ecolabel
104 criteria.

105 One of the circularity indicators most widely used or compared in the scientific literature is
106 the Material Circularity Indicator (MCI) developed by the Ellen MacArthur Foundation and Granta
107 Design (Bracquené et al., 2020; Glogic et al., 2021; Lonca et al., 2020; Mantalovas and Di Mino, 2020;
108 Muñoz et al., 2020). The MCI is based on material flow analysis and measures the maximum
109 restorative flow and the minimum linear flow of a product (Ellen MacArthur Foundation, 2015).
110 However, the MCI has been criticized for its limited application of the aggregation principles (Linder
111 et al., 2017) and for the static perspective of assumption (Pauliuk et al., 2017).

112 In LCA, two issues must be distinguished with regard to materials scarcity:

- 113 (a) materials that are wasted in a system, e.g. by emissions, dispersion, or landfill at end-of-
114 life (the issue of "waste is food" in C2C);
- 115 (b) materials that are present in a system as "materials hold-up" (related to the residence
116 time of the materials).

117 Issue (a) is dealt with in the most used indicators for materials scarcity (in ReCiPe, CML,
118 Environmental Footprint, Eco-costs). An overview of the basic philosophies behind these indicators
119 and its consequences are given in (Vogtländer et al., 2019). Issue (b) is not often dealt with in LCA,
120 since it is seldom that this issue is scope of the study.

121 In addition to the methods of circularity assessment listed above, more than fifty other
122 circularity indicators can be found in both the academic and the "grey" literature (Saidani et al., 2019),
123 all of which are applied for different purposes, scopes and usages. A variety of conceptualizations and
124 terminologies used to describe the circular economy paradigm have been reported in various reviews
125 (Corona et al., 2019; Linder et al., 2020) and research papers (Kravchenko et al., 2020; Llorente-
126 González and Vence, 2019). As many as 114 different definitions of the paradigm of the circular
127 economy were identified in a study by Kirchherr et al. (2017), leading the authors to predict a possible
128 collapse of the concept due to such broad interpretations. Achieving a unique and simple metrics to
129 serve as a general concept for monitoring the circularity of packaging and for calculating this at macro,
130 mezzo or micro scales of the analysed system is thus an essential and challenging task.

131 In this paper we address this challenge by defining simple metrics for packaging using the
132 previously defined concept for value creation in the circular economy and by combining existing
133 variables from this concept. In particular, we introduce the concept of the time that material exists in
134 the technosphere as a method to enhance of traditional LCA models, aiming at providing new insights
135 to inform scientific debate on the allocation of credits and burdens between several consecutive life
136 cycles of a product.

137 2. Methods

138 In developing new circularity indicators for assessing the circularity of packaging, we used the
139 concept and basic principles defined in two international standards for life cycle assessment: ISO
140 Standard 14040 (ISO 14040:2006(E), 2006) and ISO Standard 14044 (ISO 14044:2006(E), 2006). In
141 addition, we considered the main variables defined in the principles proposed by the Ellen MacArthur
142 Foundation for value creation in the circular economy (Ellen MacArthur Foundation, 2013). The Ellen
143 MacArthur Foundation report published in 2013 defined the following four principles for value
144 creation in the circular economy: (i) the power of the inner circle; (ii) the power of circling longer; (iii)
145 the power of cascaded use; and (iv) the power of pure circles (non-toxic, or at least easier-to-separate
146 inputs). The aim of the circularity principles is to address one-off negative effects of providing material

147 supply in a short period of time. The added value lies in the cumulative advantages of circularity
148 principles compared to the linear economic principles (Ellen MacArthur Foundation, 2013). All four
149 principles hence aim at preserving material (i.e. products, components and material) in the
150 technosphere while maximizing its energy performance over time. On this basis, the three main
151 variables we use for assessing value creation in a circular economy are energy (E), material (M) and
152 time (t).

153 Following the basic circularity principles and the three variables, we developed the equation
154 for calculation of function of the material and energy needed to produce $n + 1$ product. We illustrate
155 it with a simple case study to present the impacts of time on the traditional linear LCA model. The
156 case study shows the additional material and energy needed for a product to fulfil the same function,
157 but when the time in the technosphere is different, which is influenced solely by different time of the
158 end of life stage of the product.

159 2. Results and Discussion

160 3.1. Inclusion of time in life cycle

161 In our efforts to present the interrelation between energy, material and time in the life cycle
162 of a product, we use the adopted view of circularity and develop a new calculation metric derived
163 from these three variables.

164 We distribute three vectors (axes) in one geometric plane with the following conditions:

$$\begin{aligned} 165 \quad & \vec{x} + \vec{y} + \vec{z} = \vec{0} \\ 166 \quad & |\vec{x}| = |\vec{y}| = |\vec{z}| = 1 \\ 167 \quad & \sphericalangle \vec{x} \vec{y} = \sphericalangle \vec{x} \vec{z} = \sphericalangle \vec{y} \vec{z} = 120^\circ \end{aligned}$$

168 On axis \vec{x} we distribute the function of energy and material, which depends on the
169 production $f_{pr}(E, M)$ of the packaging. On axis \vec{y} we present the function of energy and material used
170 in the use stage $f_{us}(E, M)$ with the added constant from the previous stages. On axis \vec{z} we provide
171 the function of energy and material used at the end of life (EoL) $f_{EoL}(E, M)$ with the added constant
172 from the previous stages of the life cycle. The EoL stage here covers all aspects and impacts incurred
173 after the usage of the packaging product, i.e. its sorting, collection, treatment, recycling, the
174 reincorporation of secondary materials, and the treatment of waste (littering, incineration and
175 landfilling).

176 Figure 1 presents the relations between the three functions in closed-loop recycling, where
177 Point 1 presents the value of the production function of the 1st packaging product(s) and $f_{pr}(E, M)$
178 equals:

$$179 \quad f_{pr1}(E, M)\vec{z} = f_{pr}(E_1, M_1)\vec{z},$$

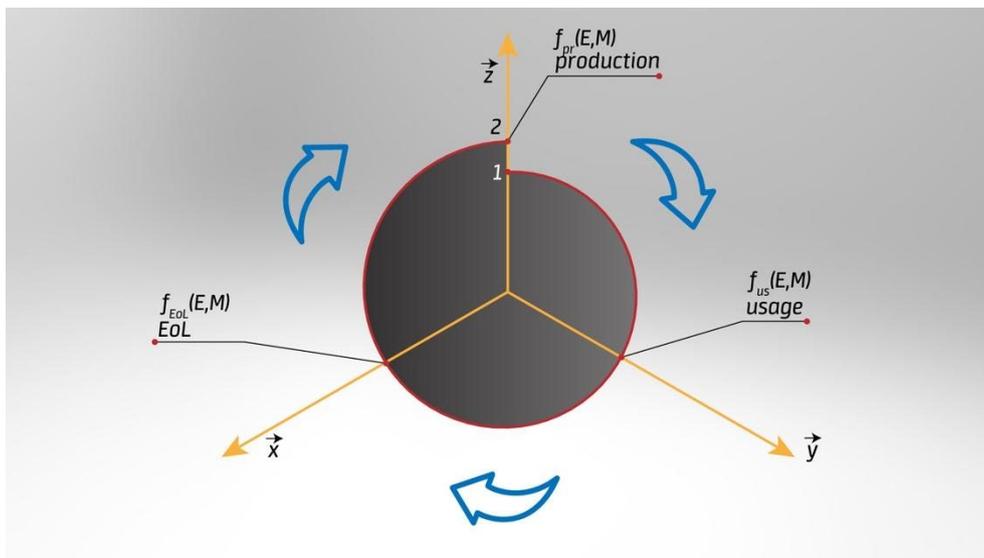
180 where E_1, M_1 are the energy and material needed for producing the 1st product. Point 2 presents
181 the value of the production function of the 2nd consecutive product and after one full life cycle of the
182 product; the value of $f_{pr}(E, M)$ equals:

$$183 \quad f_{pr2}(E, M)\vec{z} = [f_{pr}(E_2, M_2) + f_{pr}(E_1, M_1) + f_{us}(E_1, M_1) + f_{EoL}(E_1, M_1)]\vec{z}, \text{ or}$$

$$184 \quad f_{pr2}(E, M)\vec{z} = [f_{pr}(E_2, M_2) + f_1(E_1, M_1)]\vec{z}, \text{ where}$$

$$185 \quad |f_1(E_1, M_1)| = |f_{pr}(E_1, M_1)| + |f_{us}(E_1, M_1)| + |f_{EoL}(E_1, M_1)|, \text{ or}$$

186 $f_1(E_1, M_1)$ is the sum of the functions of all material and energy in the first life cycle of the
187 product. The value of $f_1(E_1, M_1)$ is higher than in all of the consecutive life cycles due to the high value
188 of $f_{pr}(E_1, M_1)$, since it is during the first cycle that the largest proportion of the material is introduced
189 in the life cycle to serve the function and functional unit determined by the producer. The extended
190 producer's responsibility of the packaging corresponds to this amount of material (Monier et al.,
191 2014). The function $f(E, M)$ in general is an ever-growing function due to the accumulation of impacts
192 arising from the additional energy and material needed for the product to keep serving the same
193 function.



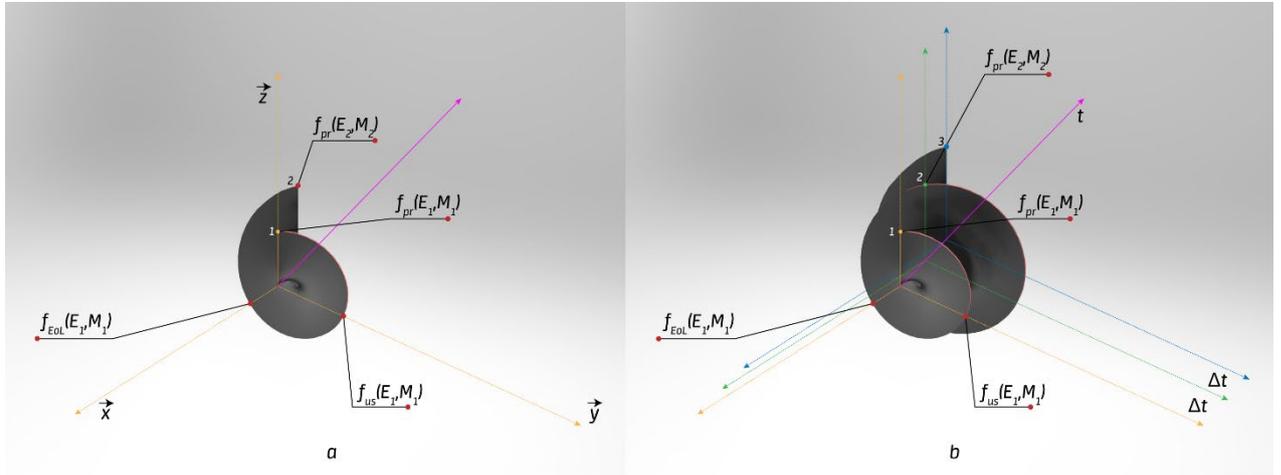
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195 *Figure 1. Ever-growing function with accumulation of impacts in a single life cycle of a product*

196 Ever since the introduction of the Life Cycle Analysis (LCA) standards (ISO 14040:2006(E),
197 2006; ISO 14041:1998(E), 1998; ISO 14044:2006(E), 2006) there has been an intense debate as to the
198 appropriate allocation of impacts amongst the consecutive life cycles. Ekvall and Finnveden (2001),
199 for example, have discussed the solution of this allocation problem in terms of subdivision, system
200 expansion, and physical causal relationships. A study by Frischknecht (2010) assessed two different
201 approaches to the end of life allocation: (i) the recycled content, or cut-off approach, and (ii) the end-
202 of-life recycling, or avoided burden approach, concluding that the cut-off approach has a stronger
203 impact on sustainability. In addition, Ekvall (2000) developed a circularity concept that takes into
204 account the influence of the market in allocation procedures in open-loop recycling, while Werner
205 and Richter (2000) discussed the implications of using an economic model of allocation. Bala et al.
206 (2015) have analysed the influence of the quality of the secondary material in the calculation of credits
207 through material recovery and substitution ratios. Allacker et al. (2014) have scrutinized five accepted
208 standards against eight criteria to assess the holistic approach, concluding that Product Environmental
209 Footprint (PEF) is one of two modelling approaches better suited for the creation of product policies.
210 In a more recent study, Allacker et al. (2017) have explained the rationale behind the PEF selection of
211 the end of life allocation modelling between the first life cycle and previous and sequent cycles. Based
212 on the present PEF Guidance (European Commission, 2018b), different allocation factors are defined
213 for different materials and all previously discussed aspects are incorporated in the equation. This
214 represents the most recent attempt of standardisation of the environmental impact on a product level
215 in the EU.

216 With an ambition to tackle and solve this long-discussed issue in the LCA scientific field, we
217 include a third important variable defined in the value creation principles of the circular economy,
218 introducing an additional axis perpendicular to the plane on which the axes \vec{x} , \vec{y} , \vec{z} are defined. This

219 variable is defined as the time axis t of the process, as presented in Figure 2a. By introducing axis t ,
 220 the three-dimensional coordination system with the function of energy and material in production,
 221 usage, and EoL, is geometrically translated in time, and the time of one life cycle is Δt (Figure 2b). Δt
 222 is constant because we use a simplified example of closed-loop recycling in the theoretically identical
 223 waste management system.



224

225 *Figure 2.* The inclusion of time in the life cycle of a product. Figure 2 a) presents a single life
 226 cycle. Figure 2 b) presents two full life cycles of a product, with point 3 representing the value of the
 227 production function of the third consecutive product, in accordance with the predefined criteria.

228 On this basis, we propose the following equation:

$$229 \quad f_3(E_3, M_3)\vec{z} = [f_{pr}(E_3, M_3) + f_1(E_1, M_1) + f_2(E_2, M_2)]\vec{z},$$

230 or we can work with the absolute values and deduct the vector \vec{z} from the equation

$$231 \quad f_3(E_3, M_3) = f_{pr}(E_3, M_3) + f_1(E_1, M_1) + f_2(E_2, M_2)$$

232 This equation can be transformed to present the " n^{th} " full cycle of the product as:

$$233 \quad f_{n+1}(E_{n+1}, M_{n+1}) = f_{pr}(E_{n+1}, M_{n+1}) + \sum_{i=1}^n f_i(E_i, M_i) \quad (\text{Equation 1})$$

234 Presenting the life cycle of the product and the three important variables (energy, material
 235 and time) for value creation in the circular economy enables the extraction of several circularity
 236 indicators, as elucidated below.

237 3.2. Circularity Indicators

238 3.2.1. Angle of circularity in the production of packaging

239 The first indicator derived from introducing time into the geometrical representation of a
 240 product life cycle relates to the quality of the material in recycling and to changes in the quality of
 241 material over time $\frac{dQ}{dt}$. The scientific literature has provided some examples of the limitations of
 242 packaging materials in closed-loop recycling. For example, Niero and Olsen (2016) have reported
 243 problems in the alloying composition of closed-loop recycled aluminium cans, limiting its recycling to
 244 a maximum of seven cycles due to the lack of magnesium and manganese for the production of the
 245 lids of such cans. (This specific case, with a maximum of seven recycling cycles, is presented in *Figure*
 246 *3.*)

247 The literature further suggests that the standard practice of mechanically recycling plastics
 248 results in quality losses due to changes in the molecular weight of the complex polymer molecule of

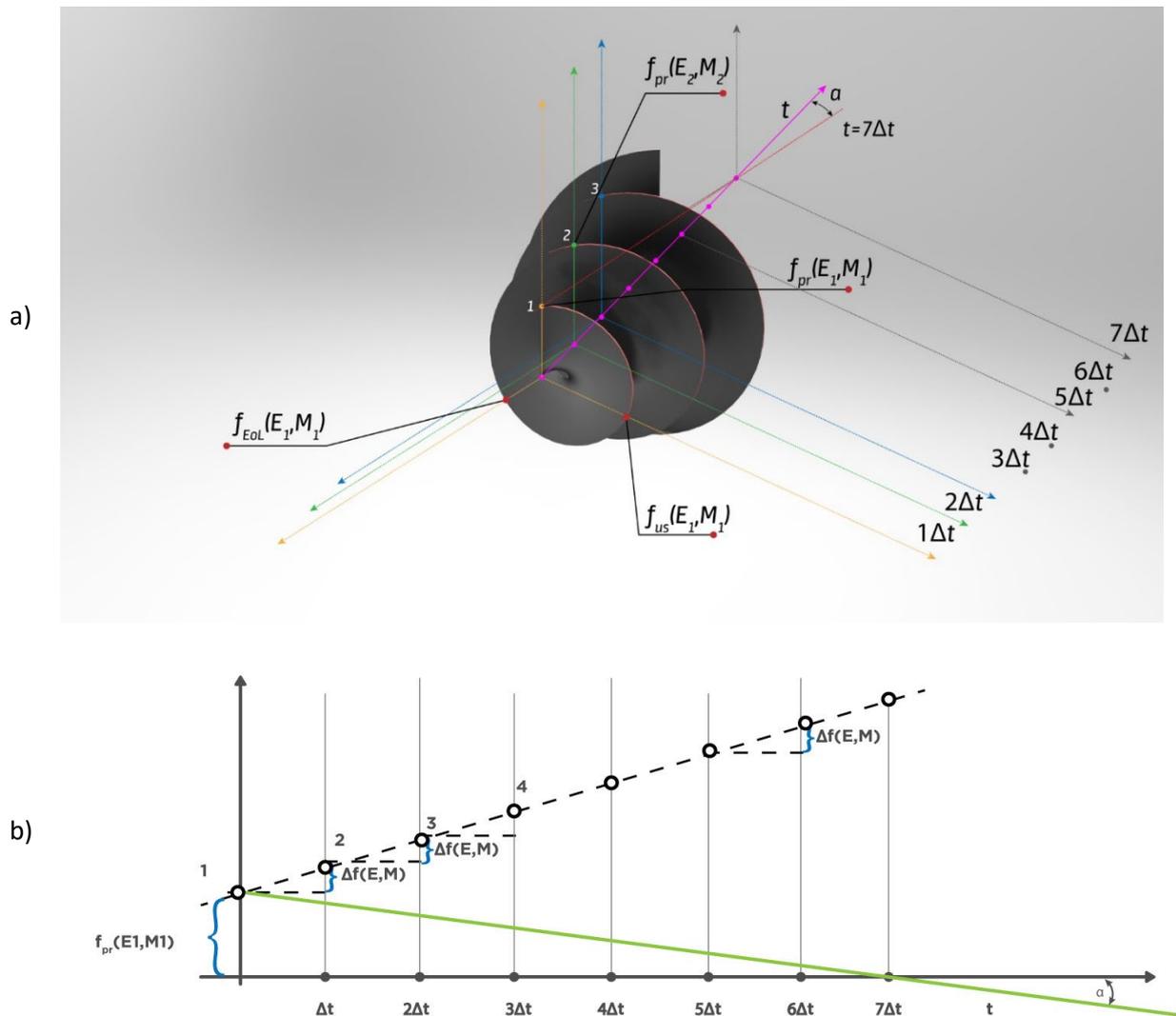
249 the secondary material (Grigore, 2017). For example, the use of secondary glass in closed-loop
 250 recycling has been shown to require a significant amount of post-treatment to meet the quality
 251 standards of the industry production chain due to different glass types and impurities (Deschamps et
 252 al., 2018), meaning laboratory testing of the mechanical, chemical, physical and optical properties of
 253 the material is needed to assess the qualitative changes of the material in closed-loop recycling.
 254

255 The possible number of life cycles in closed-loop recycling of a product's packaging before the
 256 material stops being suitable to serve the same purpose, in relation with the $f_{pr}(E_1, M_1)$ as a function
 257 of the material and energy needed for the production of the first packaging product, can be
 258 considered as an *angle of circularity in production*. The smaller the angle the more circular the
 259 packaging product. In Figure 3 the angle of circularity in production is presented as $\sphericalangle\alpha$.

260 A specific feature in the case of this indicator is that we do not add additional mass in any
 261 stage of the life cycle because of the closed-loop recycling without additional mass. This means that:

262
$$M_1, M_2, \dots, M_n = 0$$

263 for all stages of the life cycle, production, usage and EoL.



264

265 Figure 3: Angle of circularity ($\sphericalangle\alpha$) in production after seven recycling loops (Fig. 3a) and presentation in
 266 $f_{pr}(E_n, M_n)$ and t coordinate system (Fig. 3b)

267 3.2.2. Angle of circularity of the packaging life cycle

268 The scientific literature has recognized the importance of taking multiple loops into account
269 in designing life cycle assessments of packaging (Niero et al., 2016; Sazdovski et al., 2021). For
270 products such as packaging that have a short usage time in the technosphere, considering multiple
271 cycles can marginalize the problem of allocating burdens and credits amongst consecutive cycles. The
272 specific theoretical case of equation 1, where $n \rightarrow \infty$, would thus render the EoL allocation
273 insignificant in the overall life cycle assessment.

274 The specific case for this indicator is that we add additional mass in every stage of the life
275 cycle in order to keep the same function of the material for all stages of the life cycle, production,
276 usage and EoL. This means that:

$$277 M_1, M_2, \dots, M_n \neq 0$$

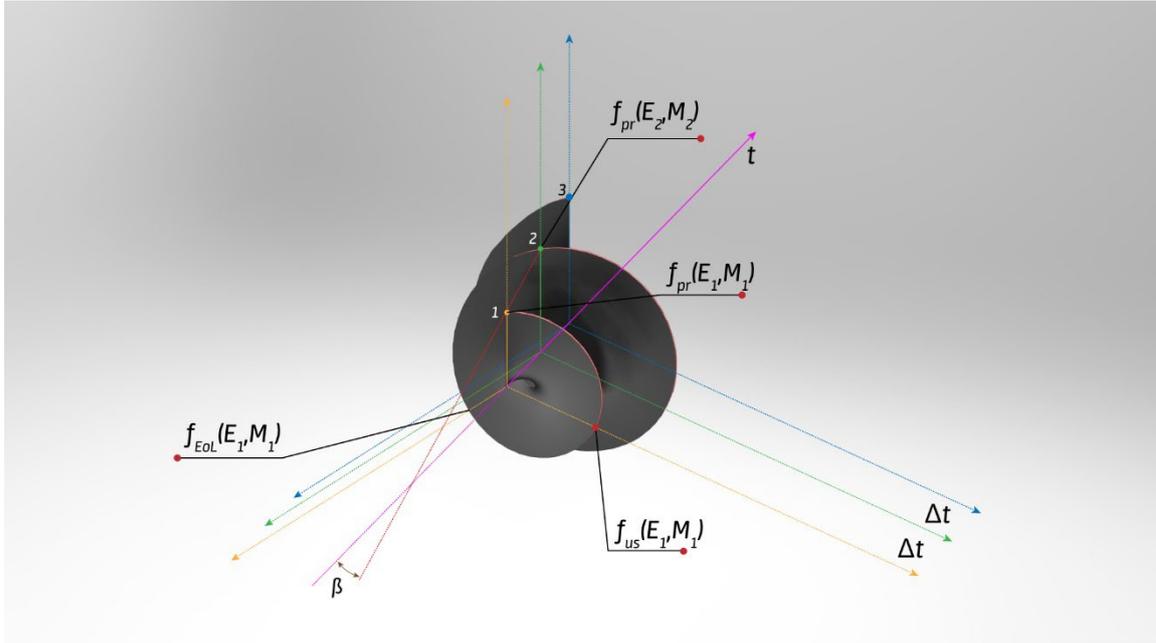
278 Packaging produced from metals and glass has a relatively high value of $f_{pr}(E_1, M_1)$ for the
279 production of the first product due to the energy-intensive production process involved. However,
280 such packaging has relatively low values of $f_{pr}(E_{n+1}, M_{n+1})$ for consecutive life cycles compared to
281 plastic and paper packaging due to the quality of the material in recycling, the recycling yields, the
282 influence of the recycled content and the small material losses incurred in the recycling process. To
283 find a breakeven point of the number of cycles for different packaging materials that serve the same
284 function and functional unit, we can use a sensitivity analysis of the number of circles of the equation
285 (1).

286 From the graphical presentation in Figure 2 that shows the three main variables that influence
287 value creation in the circular economy, we can identify one new indicator of circularity that we name
288 the *angle of circularity of the packaging*.

289 The relations between the three consecutive life cycles of the product, $n, n + 1$ and $n + 2$,
290 are represented graphically in Figure 4.

291 Here we consider a simplified case in which the material for the production of $(n + 1)^{th}$ of a
292 product has the same quality as the material used for the production of n^{th} product, serves the same
293 function, and is produced under the same conditions. We assume that the addition of the virgin
294 material is constant for the sake of simplification, although in reality the recycled part will be dirtier.

295



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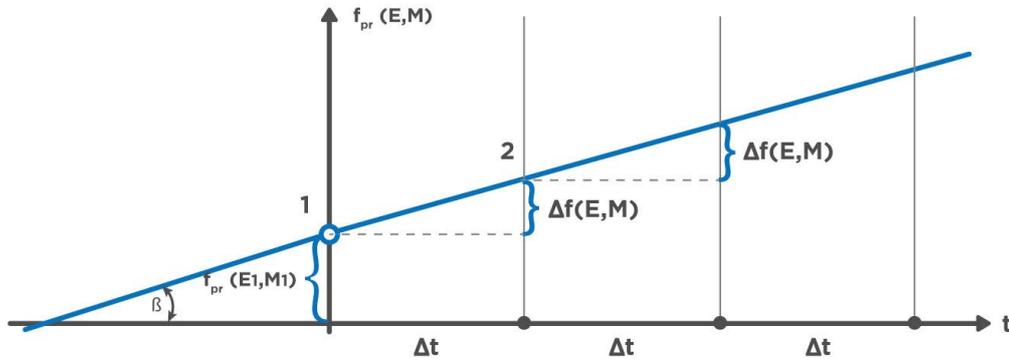
Figure 4: Angle of circularity ($\alpha\beta$)

298

299

300

We can define a straight line that connects the value of $f_{pr}(E_n, M_n)$ and the value of $f_{pr}(E_{n+1}, M_{n+1})$ presented as Point 1 and Point 2 in Figure 4. Figure 5 presents the dependency between the $f_{pr}(E, M)$ and time in the coordination system.



301

302

Figure 5: Angle of circularity in $f_{pr}(E_n, M_n)$ and t coordinate system

303

304

305

306

Angle β is defined as the point of intersection of the previously defined line and the axis of time t . This angle can be defined as an angle of circularity and helps us to observe the circularity of the material after $n + 1$ life cycles. The smaller the angle of circularity, the better the circular properties of the material. The value of $\Delta f(E, M)$ equals:

307

$$\Delta f(E, M) = f_{us}(E_n, M_n) + f_{EoL}(E_n, M_n) + f_{pr}(E_{n+1}, M_{n+1})$$

308

309

and $\Delta f(E, M) = f_n(E_n, M_n)$ when equal energy and material is consumed for the production of each consecutive cycle (n and $n+1$).

310 On the basis of this conclusion we can say that the angle of circularity depends on the function
311 of the energy and material in each cycle and the time of the life cycle in the technosphere.

312 In an ideal situation where the material introduced in the technosphere stays there
313 indefinitely and is reused in the consecutive life cycles, the value of the angle of circularity and the
314 angle of circularity of production is equal to zero. This means that no additional material is added for
315 serving the same function. No additional energy is added in the system under the analysis either. In
316 reality, the angles however depend on the qualitative properties of the material that is recycled, as
317 well as on the energy needed for the recycling as essential variables. Finally, the angles will depend
318 on the real time of the life cycle, being in particularly impacted by the t of EoL.

319 The influence of time on angle of circularity is explained in the following section.

320 3.3. The relativity of time in the technosphere for packaging products

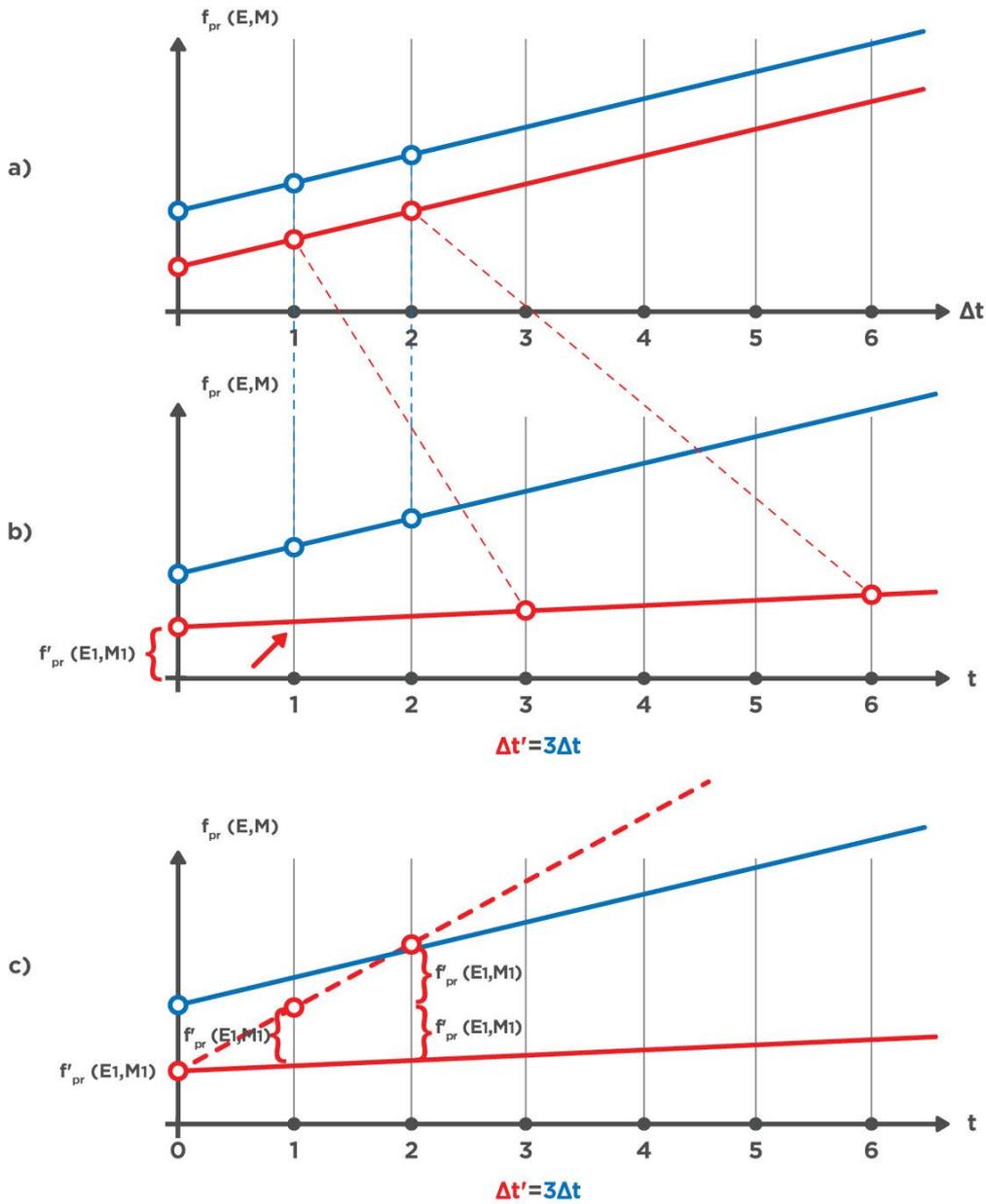
321 Time is a very important variable in the life cycle of packaging. The packaging production
322 process has become simpler and faster due to the continuously expanding market and increasing
323 demand. The maximal time of the usage phase of the life cycle is linked to the time that a product
324 (beverage or food) is fit for consumption, hence the usage time does not depend on the packaging
325 material, but on the packaged good, meaning the time of usage will not change if you replace the
326 packaging for a different material with the same function.

327 The biggest difference occurs in the end of life stage of packaging materials. This difference
328 comes from the varying efficiency of waste management systems, the efficiency of sorting systems,
329 the efficiency of the recycling of the material, the recycling yields, and the location of the recycling
330 facility. The last factor (location) is related to the time needed for transportation of the sorted and
331 recycled material. The following equation represents the time of one life cycle of the packaging:

$$332 \Delta t = t_{pr} + t_{us} + t_{EoL}$$

333 The effect of including the real time of the life cycle in the production of packaging is
334 presented in Figure 6.

335



336

337 *Figure 6.* Relativity of time in the life cycle of two packaging systems. Difference in the function of
 338 energy and material from life cycle (6a), and from time (6b and 6c) of two systems. In the first
 339 system $\Delta t = t$ (represented in blue) and in the other system $\Delta t' = 3t$ (represented in red).
 340

341 Figure 6a illustrates two different packaging systems that use the same material but with
 342 different functions of energy and material for the production of the first packaging $f_{pr}(E_1, M_1)$. The
 343 packaging represented in blue finalizes one full life cycle exactly on time for the production of new
 344 packaging from the secondary material. The life cycle of the packaging represented in red has $\Delta t'$
 345 three times longer due to the lower efficiency of the waste management system (Figure 6b) and
 346 production of the first packaging $f'_{pr}(E_1, M_1)$.

347 Even though there is no difference in the relation between the $f(E, M)$ and the life cycle, a
 348 geometrical translation in time, as presented in Figure 6b (see the red arrow in the graph) shows a
 349 lack of packaging material for reproduction because the secondary material did not finalize its life

350 cycle and therefore new additional material for the production of the packaging of the following cycle
 351 needs to be introduced into the system to serve the same function and functional unit (Figure 6c).
 352 The amount of this additional $f(E, M)$ is proportional to the relation between the time of the life
 353 cycle and the time (t_{need}) needed for the secondary material to be prepared for a new cycle (the need
 354 for a new product) of the life cycle of the packaging.

355 Bearing all this in mind, we can transform Equation 1 in order to present the function of the
 356 material and energy for the production of the packaging after n number of circles:

$$357 \quad f_{n+1}(E_{n+1}, M_{n+1}) = f_{pr}(E_{n+1}, M_{n+1}) + \sum_{i=1}^n f_i(E_i, M_i) + \left(\frac{\Delta t}{t_{need}} - 1\right) f'_1(E_1, M_1)$$

358 *(Equation 2)*

359 *where: $\Delta t, t_{need} \in N \wedge \Delta t \geq t_{need}$*

360 Also, $\Delta t, t_{need}$ are expressed in the same time unit appropriate for the need for the
 361 production stage of packaging.

362 Where $\left(\frac{\Delta t}{t_{need}} - 1\right) f'_{pr}(E_1, M_1)$ in the Equation 2 is the additional number of packaging
 363 products needed for balancing the system until the end of every life cycle. Sazdovski et al. (2021) have
 364 reported that the discrepancy between the exported and imported material of the packaging is
 365 related to the fact that most plastic packaging is transported from developed countries to developing
 366 countries during the EoL. This can create a significant discrepancy in the time of the life cycle, even
 367 though it does not contribute to the circularity of the product. In addition, the time of EoL of some
 368 materials might be significantly longer than the time for production and the time of usage in the life
 369 cycle. This discrepancy in times is not taken into consideration in linear LCA calculations. As presented
 370 in *equation 2*, this discrepancy results in adding additional material and energy into the technosphere
 371 for serving the same functional unit. The longer the life cycle of the material, the more additional
 372 packaging material is needed.

373 The different industries producing fast moving products generally do not disclose the average
 374 time of the EoL of their products, with the aluminium industry being an exception by clamming that
 375 an aluminium can is possible to be reproduced in 60 days (3BL CSRwire, 2010; Belinda, 2006;
 376 EarthShare, 2008; Island Return It Recycling Depot, 2013; Leahy, 2019; Marck Recycling, 2015; Novelis,
 377 2021; Recyclebank, 2012; Reyes, 2010; Richmond Steel Recycling, 2019; The Aluminum Association,
 378 2014; Think Cans, 2017). The industry considers this as a circular property of the aluminium can. Not
 379 having information about comparative products serving similar function, we will conduct a simple
 380 comparison using two equal aluminium cans. The cans are produced under equal conditions, but they
 381 follow two different EoL scenarios: one is two times longer than the other. Therefore, we can
 382 transform the *Equation 2* to calculate the function of E and M for development of a daily production
 383 of “ m ” aluminium cans, with a life cycle of 60 days.

$$384 \quad mf_{n+1}(E_{n+1}, M_{n+1})$$

$$385 \quad = mf_{pr}(E_{n+1}, M_{n+1}) + m \sum_{i=1}^n f_i(E_i, M_i) + [(t_{pr} + t_{us}$$

$$386 \quad + 60days)/1day - 1] mf'_1(E_1, M_1)$$

387 *(Equation 3.)*

388 We use the same equation to calculate the function of E and M of the same production
 389 process for the aluminium cans after their use end up in a EoL that have the same function
 390 $f_{EoL}(E_n, M_n)$ but two times longer duration ($t_{EoL} = 120days$). Having in mind that the production

391 process is the same and assuming that the usage is the same, we can transform the equation 2 for
392 this process.

$$\begin{aligned} 393 \quad & mf''_{n+1}(E_{n+1}, M_{n+1}) \\ 394 \quad & = mf_{pr}(E_{n+1}, M_{n+1}) + m \sum_{i=1}^n f_i(E_i, M_i) + [(t_{pr} + t_{us} \\ 395 \quad & + 120days)/1day - 1] mf'_{pr}(E_1, M_1) \\ 396 \quad & \text{(Equation 4.)} \end{aligned}$$

397 Simple deduction between the Equation 4 and the Equation 3 will show us the additional
398 material needed to serve the same function. The need for the additional material is only influenced
399 by the length of the EoL.

$$400 \quad mf''_{n+1}(E_{n+1}, M_{n+1}) - mf_{n+1}(E_{n+1}, M_{n+1}) = 60mf'_{pr}(E_1, M_1)$$

401 The result is 60 daily productions of the primary material $f'_{pr}(E_1, M_1)$ that are needed to be
402 added in the technosphere to serve the same function. The linear LCAs do not calculate this
403 additional material. This simple example illustrates the impact of time on life cycle, showing the
404 importance of considering time when conducting life cycle analysis of packaging.

405 The findings provided and discussed in this section suggest that when considering the time
406 variable in the life cycle of a packaging product, the key difference that appears between packaging
407 materials is related to the time of the material's EoL. When comparing the life cycles of packaging
408 made from different materials, therefore, the time of the life cycle can be equated to the time of the
409 EoL:

$$410 \quad \Delta t \cong t_{EoL} \cong t_{EoLmax}$$

411 In order to make a comparison of different packaging materials using the previously
412 developed circularity indicators, we would need to define the maximal time of the EoL to be
413 considered for all materials performing the same function. This would enable a comparison of all the
414 materials under the same circumstances. This approach could bring improvements in the time
415 efficiency of waste management systems and the sorting of packaging. Finally, this approach could
416 also stimulate the creation of a local market for secondary materials.

417 3. Conclusions

418 The circular economy is a relatively new concept introduced to overcome the limitations of
419 the traditional linear economy model that has proved to be inefficient in terms of material use and
420 environmental protection. A growing body of scientific literature aims at providing adequate metrics
421 in order to improve comparison between the two economic paradigms or systems serving the same
422 function. One of the key limitations of the circular economy concept is the lack of a unique definition
423 and of commonly accepted indicators for circularity. Moreover, there is a plethora of different
424 approaches utilizing or introducing new variables that are not defined in previously developed
425 conceptual frameworks of the circular economy.

426 Accordingly, this paper starts from the widely accepted and referenced definition of circularity
427 in order to ascertain which key variables need to be considered. On this basis, our paper introduces
428 for the first time a combination of the three variables that define the value creation principles
429 developed by the Ellen MacArthur Foundation: material, energy and time. When the materials hold-
430 up in a system is regarded as important, the time as a third variable must be taken into account.. This
431 is especially the case for products that have a relatively short usage time in the technosphere

432 compared to the time needed for recycling and waste management. Packaging is a perfect example
433 of such a product.

434 We developed a formula for calculating the so-called " $n + 1$ " product that includes the time
435 necessary for obtaining the secondary material needed for its production. According to this formula,
436 we need to consider that the production of additional packaging products and the quantity of these
437 products depend on the time needed for recycling in the waste management system. The t_{EoL} for
438 different packaging materials varies, which makes LCA in circular economy difficult to compare with
439 the tradition linear LCA method.

440 This is similar to the stocking of packaging in reusable packaging systems that serves the
441 function of the system ensuring packaging availability. Traditionally, this phenomenon has not been
442 taken into account when developing attributional LCAs between systems that serve the same
443 function.

444 The representation of the life cycle of packaging proposed here addresses the problem widely
445 discussed in scientific circles related to the allocation of credits and burdens between several
446 consecutive life cycles. We have demonstrated this issue for the specific case in which $n \rightarrow \infty$ and
447 applied our formula to calculate the " $(n + 1)^{th}$ " product.

448 This study showed how to improve the LCA practice by including circularity principles. While
449 impacts from the usage of energy and material is studied in the linear LCA model, here we exposed
450 additional considerations by adding time variable in the calculation. This is particularly obvious for
451 fast-moving goods, such as packaging. For comparative LCAs, introduction of time as a variable can
452 also serve for development of the circularity indicators. The presented circularity indicators can be
453 used for the following two purposes: i) for assessing the circularity of packaging in production; and ii)
454 for assessing the circularity of packaging during the lifetime of packaging materials. Future research
455 should assess the application of the developed indicators in open-loop and closed-loop recycling.

456

457

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Bibliography

- 3BL CSRwire, 2010. Novelis Recycles a Record 40 Billion Beverage Cans in 2009 [WWW Document]. URL https://www.csrwire.com/press_releases/29345-novelis-recycles-a-record-40-billion-beverage-cans-in-2009
- Allacker, K., Mathieux, F., Manfredi, S., Pelletier, N., De Camillis, C., Ardente, F., Pant, R., 2014. Allocation solutions for secondary material production and end of life recovery: Proposals for product policy initiatives. *Resour. Conserv. Recycl.* 88, 1–12. <https://doi.org/10.1016/j.resconrec.2014.03.016>
- Allacker, K., Mathieux, F., Pennington, D., Pant, R., 2017. The search for an appropriate end-of-life formula for the purpose of the European Commission Environmental Footprint initiative. *Int. J. Life Cycle Assess.* 22, 1441–1458. <https://doi.org/10.1007/s11367-016-1244-0>
- Bala, A., Raugei, M., Fullana-i-Palmer, P., 2015. Introducing a new method for calculating the environmental credits of end-of-life material recovery in attributional LCA. *Int. J. Life Cycle Assess.* 20, 645–654. <https://doi.org/10.1007/s11367-015-0861-3>
- Battle-Bayer, L., Bala, A., Lemaire, E., Albertí, J., García-Herrero, I., Aldaco, R., Fullana-i-Palmer, P., 2019. An energy- and nutrient-corrected functional unit to compare LCAs of diets. *Sci. Total Environ.* 671, 175–179. <https://doi.org/10.1016/j.scitotenv.2019.03.332>
- Belinda, M., 2006. Analysis of the Recycling Method for Aluminum Soda Cans.
- Bracquené, E., Dewulf, W., Duflou, J.R., 2020. Measuring the performance of more circular complex product supply chains. *Resour. Conserv. Recycl.* 154, 104608. <https://doi.org/10.1016/j.resconrec.2019.104608>
- Cordella, M., Alfieri, F., Sanfelix, J., Donatello, S., Kaps, R., Wolf, O., 2020. Improving material efficiency in the life cycle of products: a review of EU Ecolabel criteria. *Int. J. Life Cycle Assess.* 25, 921–935. <https://doi.org/10.1007/s11367-019-01608-8>
- Corona, B., Shen, L., Reike, D., Rosales Carreón, J., Worrell, E., 2019. Towards sustainable development through the circular economy—A review and critical assessment on current circularity metrics. *Resour. Conserv. Recycl.* 151, 104498. <https://doi.org/10.1016/j.resconrec.2019.104498>
- Council of the European Union, 1985. Council Directive 85/339/EEC of 27 June 1985 on containers of liquids for human consumption. European Commission.
- Deschamps, J., Simon, B., Tagnit-Hamou, A., Amor, B., 2018. Is open-loop recycling the lowest preference in a circular economy? Answering through LCA of glass powder in concrete. *J. Clean. Prod.* 185, 14–22. <https://doi.org/10.1016/j.jclepro.2018.03.021>
- Di Maio, F., Rem, P.C., Baldé, K., Polder, M., 2017. Measuring resource efficiency and circular economy: A market value approach. *Resour. Conserv. Recycl.* 122, 163–171. <https://doi.org/10.1016/j.resconrec.2017.02.009>
- EarthShare, 2008. Green Quiz: Recycling Aluminum Cans [WWW Document]. URL <https://www.earthshare.org/earthshare-quiz-november-7-2008/>
- Ekvall, T., 2000. A market-based approach to allocation at open-loop recycling. *Resour. Conserv. Recycl.* 29, 91–109. [https://doi.org/10.1016/S0921-3449\(99\)00057-9](https://doi.org/10.1016/S0921-3449(99)00057-9)
- Ekvall, T., Finnveden, G., 2001. Allocation in ISO 14041 - a critical review. *J. Clean. Prod.* 9, 197–208. [https://doi.org/10.1016/S0959-6526\(00\)00052-4](https://doi.org/10.1016/S0959-6526(00)00052-4)

- Ellen MacArthur Foundation, 2015. CIRCULARITY INDICATORS: An Approach to Measuring Circularity: METHODOLOGY, Ellen MacArthur Foundation (EMF). https://doi.org/https://www.ellenmacarthurfoundation.org/assets/downloads/insight/Circularity-Indicators_Methodology_May2015.pdf
- Ellen MacArthur Foundation, 2013. Towards the circular economy; Economic and business rationale for an accelerated transition.
- European Commission, 2020. Roadmap - New Circular Economy Action Plan. Brussels.
- European Commission, 2019a. REPORT FROM THE COMMISSION TO THE EUROPEAN PARLIAMENT, THE COUNCIL, THE EUROPEAN ECONOMIC AND SOCIAL COMMITTEE AND THE COMMITTEE OF THE REGIONS on the implementation of the Circular Economy Action Plan, European Commission. Brussels.
- European Commission, 2019b. DIRECTIVE (EU) 2019/904 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL of 5 June 2019 on the reduction of the impact of certain plastic products on the environment, Official Journal of the European Union. European Commission, Brussels, Belgium.
- European Commission, 2018a. DIRECTIVE (EU) 2018/852 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL of 30 May 2018 amending Directive 94/62/EC on packaging and packaging waste (Text with EEA relevance), Official Journal of the European Union.
- European Commission, 2018b. Product Environmental Footprint Category Rules Guidance. PEFCR Guid. Doc. 238.
- European Union, 1994. European Parliament and Council Directive 94/62/EC on Packaging and Packaging Waste, Official Journal of the European Communities No L 365/10. <https://doi.org/10.1038/sj.bdj.4811054>
- Eurostat, 2018. Packaging waste statistics [WWW Document]. Packag. waste Stat. URL https://ec.europa.eu/eurostat/statistics-explained/index.php/Packaging_waste_statistics#Waste_generation_by_packaging_material (accessed 3.3.21).
- Ferrara, C., De Feo, G., 2018. Life cycle assessment application to the wine sector: A critical review. *Sustain.* 10. <https://doi.org/10.3390/su10020395>
- Frischknecht, R., 2010. LCI modelling approaches applied on recycling of materials in view of environmental sustainability, risk perception and eco-efficiency. *Int. J. Life Cycle Assess.* 15, 666–671. <https://doi.org/10.1007/s11367-010-0201-6>
- García-Barragán, J.F., Eyckmans, J., Rousseau, S., 2019. Defining and Measuring the Circular Economy: A Mathematical Approach. *Ecol. Econ.* 157, 369–372. <https://doi.org/10.1016/j.ecolecon.2018.12.003>
- Geng, Y., Fu, J., Sarkis, J., Xue, B., 2012. Towards a national circular economy indicator system in China: An evaluation and critical analysis. *J. Clean. Prod.* 23, 216–224. <https://doi.org/10.1016/j.jclepro.2011.07.005>
- Glogic, E., Sonnemann, G., Young, S.B., 2021. Environmental trade-offs of downcycling in circular economy: Combining life cycle assessment and material circularity indicator to inform circularity strategies for alkaline batteries. *Sustain.* 13, 1–12. <https://doi.org/10.3390/su13031040>
- Gomes, T.S., Visconte, L.L.Y., Pacheco, E.B.A. V., 2019. Life Cycle Assessment of Polyethylene Terephthalate Packaging: An Overview. *J. Polym. Environ.* 27, 533–548.

<https://doi.org/10.1007/s10924-019-01375-5>

- Grigore, M.E., 2017. Methods of recycling, properties and applications of recycled thermoplastic polymers. *Recycling* 2, 1–11. <https://doi.org/10.3390/recycling2040024>
- Hey, C., 2005. *EU Environmental Policy Handbook; A Critical Analysis of EU Environmental Legislation*. Brussels, Belgium.
- Island Return It Recycling Depot, 2013. Pop Can; From Recycling Bin to Shelf in 60 Days [WWW Document]. URL <https://islandreturnit.com/2013/12/19/pop-can-from-recycling-bin-to-shelf-in-60-days/>
- ISO 14040:2006(E), 2006. *Environmental management — Life cycle assessment — Principles and framework*. Iso 140402006 2006.
- ISO 14041:1998(E), 1998. *Environmental management — Life cycle assessment — Goal and scope definition and inventory analysis*. *Environ. Manag. - Life Cycle Assess. - Goal Scope Defin. Invent. Anal.* 1998.
- ISO 14044:2006(E), 2006. *Environmental management — Life cycle assessment — Requirements and guidelines*.
- Janik, A., 2017. The analysis of existing indicators, methods and tools for measuring circularity in the context of their future development. *Int. Multidiscip. Sci. GeoConference Surv. Geol. Min. Ecol. Manag. SGEM* 17, 61–68. <https://doi.org/10.5593/sgem2017/53/S21.006>
- Kirchherr, J., Reike, D., Hekkert, M., 2017. Conceptualizing the circular economy: An analysis of 114 definitions. *Resour. Conserv. Recycl.* 127, 221–232. <https://doi.org/10.1016/j.resconrec.2017.09.005>
- Kravchenko, M., McAlloone, T.C., Pigosso, D.C.A., 2020. To what extent do circular economy indicators capture sustainability? *Procedia CIRP* 90, 31–36. <https://doi.org/10.1016/j.procir.2020.02.118>
- Leahy, M., 2019. Aluminum Recycling in the Circular Economy [WWW Document]. URL <https://www.rubicon.com/blog/aluminum-recycling/>
- Linder, M., Boyer, R.H.W., Dahllöf, L., Vanacore, E., Hunka, A., 2020. Product-level inherent circularity and its relationship to environmental impact. *J. Clean. Prod.* 260. <https://doi.org/10.1016/j.jclepro.2020.121096>
- Linder, M., Sarasini, S., van Loon, P., 2017. A Metric for Quantifying Product-Level Circularity. *J. Ind. Ecol.* 21, 545–558. <https://doi.org/10.1111/jiec.12552>
- Llorente-González, L.J., Vence, X., 2019. Decoupling or “decaffing”? The underlying conceptualization of circular economy in the European union monitoring framework. *Sustain.* 11. <https://doi.org/10.3390/su11184898>
- Lonca, G., Lesage, P., Majeau-Bettez, G., Bernard, S., Margni, M., 2020. Assessing scaling effects of circular economy strategies: A case study on plastic bottle closed-loop recycling in the USA PET market. *Resour. Conserv. Recycl.* 162, 105013. <https://doi.org/10.1016/j.resconrec.2020.105013>
- Mantalovas, K., Di Mino, G., 2020. Integrating circularity in the sustainability assessment of asphalt mixtures. *Sustain.* 12, 9–12. <https://doi.org/10.3390/su12020594>
- Marck Recycling, 2015. *Facts About Recycling Aluminum* [WWW Document]. URL <https://www.marck.net/facts-about-recycling-aluminum/>

- Molina-Besch, K., Wikström, F., Williams, H., 2019. The environmental impact of packaging in food supply chains—does life cycle assessment of food provide the full picture? *Int. J. Life Cycle Assess.* 24, 37–50. <https://doi.org/10.1007/s11367-018-1500-6>
- Monier, M., Hestin, M., Cavé, J., Laureysens, I., Watkins, E., Reisinger, H., Porsch, L., 2014. Development of guidance on Extended Producer Responsibility (EPR), European Commission.
- Muñoz, V.G., Muneta, L.M., Carrasco-Gallego, R., Marquez, J. de J., Hidalgo-Carvajal, D., 2020. Evaluation of the circularity of recycled pla filaments for 3D printers. *Appl. Sci.* 10, 1–12. <https://doi.org/10.3390/app10248967>
- Niero, M., Negrelli, A.J., Hoffmeyer, S.B., Olsen, S.I., Birkved, M., 2016. Closing the loop for aluminum cans: Life Cycle Assessment of progression in Cradle-to-Cradle certification levels. *J. Clean. Prod.* 126, 352–362. <https://doi.org/10.1016/j.jclepro.2016.02.122>
- Niero, M., Olsen, S.I., 2016. Circular economy: To be or not to be in a closed product loop? A Life Cycle Assessment of aluminium cans with inclusion of alloying elements. *Resour. Conserv. Recycl.* 114, 18–31. <https://doi.org/10.1016/j.resconrec.2016.06.023>
- Novelis, 2021. The Life of a Can – Back on the Shelf in About 60 Days [WWW Document]. URL <https://www.novelis.com/the-life-of-a-can/>
- Pauliuk, S., Kondo, Y., Nakamura, S., Nakajima, K., 2017. Regional distribution and losses of end-of-life steel throughout multiple product life cycles—Insights from the global multiregional MaTrace model. *Resour. Conserv. Recycl.* 116, 84–93. <https://doi.org/10.1016/j.resconrec.2016.09.029>
- Recyclebank, R., 2012. Benefits of Aluminum Can Recycling [WWW Document]. URL <https://livegreen.recyclebank.com/benefits-of-aluminum-can-recycling>
- Reyes, H., 2010. Recycling Cans Does More Than Help Our Environment [WWW Document]. URL <https://www.sarecycling.com/uncategorized/recycling-cans-does-more-than-help-our-environment/>
- Richmond Steel Recycling, 2019. The circular economy: Do aluminum cans really last forever? [WWW Document]. URL <https://richmondsteel.ca/aluminum-cans-recycling/>
- Saidani, M., Yannou, B., Leroy, Y., Cluzel, F., Kendall, A., 2019. A taxonomy of circular economy indicators. *J. Clean. Prod.* 207, 542–559. <https://doi.org/10.1016/j.jclepro.2018.10.014>
- Sazdovski, I., Bala, A., Fullana-i-Palmer, P., 2021. Linking LCA literature with circular economy value creation: A review on beverage packaging. *Sci. Total Environ.* 771, 145322. <https://doi.org/10.1016/j.scitotenv.2021.145322>
- The Aluminum Association, 2014. Aluminum Can Recycling Holds at Historically High Levels [WWW Document]. URL <https://www.aluminum.org/news/aluminum-can-recycling-holds-historically-high-levels>
- Think Cans, 2017. 10 aluminium facts [WWW Document]. URL <https://thinkcans.net/aluminium/10-aluminium-facts>
- Vogtländer, J., Peck, D., Kurowicka, D., 2019. The eco-costs of material scarcity, a resource indicator for LCA, derived from a statistical analysis on excessive price peaks. *Sustain.* 11. <https://doi.org/10.3390/su11082446>
- Werner, F., Richter, K., 2000. Economic allocation in LCA: A case study about aluminium window frames. *Int. J. Life Cycle Assess.* 5, 79–83. <https://doi.org/10.1007/BF02979727>

