1	Flood risk assessment in urban catchments using multiple
2	regression analysis
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6	
7	Abstract
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9	Flood assessment in urban catchments is usually addressed through the combination of
10	Geographic Information Systems (GIS) and stormwater models. However, the coupled use of
11	these tools involves a level of detail in terms of hydrological modelling which can be beyond
12	the scope of overall flood management planning strategies. This research consists of the
13	development of a methodology based on Multiple Regression Analysis (MRA) to assess flood
14	risk in urban catchments according to their morphologic characteristics and the geometrical
15	and topological arrangement of the drainage networks into which they flow. Stormwater
16	models were replaced by a combination of Multiple Linear Regression (MLR), Multiple Non-
17	Linear Regression (MNLR) and Multiple Binary Logistic Regression (MBLR), which enabled

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⁴ Associate Professor, GITECO Research Group, Universidad de Cantabria, Av. de los Castros 44, 39005, Santander, Spain (corresponding author). E-mail: <u>rodrighj@unican.es</u> identifying influential parameters in the maximum runoff rates generated in urban catchments, modelling the magnitude of peak flows across them and estimating flood risk in the nodes of sewer networks, respectively. The results obtained through a real urban catchment located in Espoo (Finland), demonstrated the usefulness of the proposed methodology to provide an accurate replication of flood occurrence in urban catchments due to intense storm events favored by Climate Change, information that can be used to plan and design preventive drainage strategies.

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26 Keywords

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Flood Risk Management; Geographic Information System; Multiple Regression Analysis;
Urban Hydrology

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31 Introduction

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The interactive effects of urbanization and Climate Change are one of the most important 33 challenges with which human beings will have to deal as a collective in the future (Hoornweg 34 et al. 2011). Their combination inflicts a particularly forceful impact on stormwater drainage 35 36 in urban catchments. Growing urbanization contributes to increasing runoff volume and 37 accelerating the time until peak flow occurs, whilst Climate Change is expected to result in an intensification of the hydrological cycle, which might lead to more violent rainfall events 38 39 (Huntington 2006). In the end, this might result in localized floods along the sewer networks 40 used to drain urban catchments, which are incapable of conveying such large amounts of water. 41 Urban floods have important consequences in physical, economic, environmental and 42 social terms (Tingsanchali 2012). These effects have traditionally been divided into tangible 43 and intangible, depending on whether they can be monetized or not (Smith and Ward 1998). 44 Examples of the former include damage to property or loss of profits, whilst the latter stand for aspects like loss of life or negative impacts on the well-being and environment. The potential 45 46 impacts of urban floods can also be classified into direct and indirect, according to their 47 spatiotemporal scale. Hence, direct damage is related to any loss caused by the immediate interaction of floods with human beings, properties and the environment, whilst indirect effects 48 concern those which are beyond the limits of the flooded area (Hammond et al. 2015). All these 49 potential consequences are expected to become more severe and frequent due to the 50 combination of Climate Change with high density of population and large impervious areas 51 52 (Huong and Pathirana 2013).

53 However, despite the increasingly important threat posed by this phenomenon, flood 54 management keeps being usually addressed in literature through stormwater models, which are often used in combination with Geographic Information Systems (GIS). Knebl et al. (2005) 55 integrated them in a study to highlight the importance of the degree of imperviousness in urban 56 catchments, which resulted in a decrease in infiltration capacity of the terrain and increased 57 flood risk. Barco et al. (2008) coupled both components to prove that imperviousness and 58 depression storage were the most influential factors in the generation of flow rates in a large 59 60 urban catchment located in Southern California. Dongquan et al. (2009) combined stormwater models with GIS based on elevation-related data such as flow direction, raster-vector 61 conversion or catchment division to automate the process of rainfall-runoff modelling. Guan 62 et al. (2015) merged both tools to determine the increase in peak flow and runoff volume caused 63 by urbanization in a catchment located in Espoo (Finland) and proposed different alternative 64 65 drainage techniques to mitigate these effects. Eshtawi et al. (2016) coupled three existing models (SWAT, MODFLOW and MT3DMS) to quantify surface-groundwater interactions in 66 67 increasingly developed areas, demonstrating the strength of using integrated hydrologic models 68 in the sustainable urban water planning process. Jato-Espino et al. (2016a) presented a GIS-69 based stormwater modelling approach that demonstrated the potential of permeable pavements 70 and green roofs to attenuate floods in urban catchments in comparison with conventional sewer 71 networks. Beck et al. (2017) described a semi-distributed approach to estimate runoff 72 reductions (TELR) to inform stormwater management decisions, including a series of algorithms for rainfall-runoff transformation and routing and specifications to implement Best 73 Management Practices (BMPs). Hanington et al. (2017) developed and calibrated a fine-scaled 74 75 quasi-2D hydro-dynamic model (IWRM-LXQ) for interprovincial water resource planning and 76 management, arguing that their approach was especially suitable for assessing hydraulically 77 complex systems at a provincial or district level.

78 The use of tools as those mentioned above is complex and entails a substantial time investment, aspects which are in conflict with the simplicity and promptness required by 79 administrative and public entities to design flood management planning strategies (Ashley et 80 al. 2007). Stormwater models are especially demanding in terms of characterization and 81 82 simulation to be used by general public or non-modelling planners (Elliott and Trowsdale 83 2007), since they involve making decisions related to infiltration and routing processes and calibrating the parameters that influence the transformation of rainfall to runoff. In contrast, 84 85 the GIS-related tasks required for creating the input data into these models to run stormwater which mainly concern catchment delineation and the determination of 86 simulations. subcatchment imperviousness and average slope, are easy to compute using basic editing and 87 88 statistical tools (Jato-Espino et al. 2016a). Furthermore, GIS are widely implemented systems 89 for multiple purposes related to the management of all kinds of spatial data, so that using them 90 for flood management does not involve an innovation with respect to common resources and 91 practices.

92 As a result of these considerations, the objective of this research was to develop a 93 methodology for flood risk assessment omitting the use of stormwater models. This was 94 achieved through the integration of Multiple Linear Regression (MLR), Multiple Non-Linear 95 Regression (MNLR) and Multiple Binary Logistic Regression (MBLR), which were combined 96 to play the role of stormwater models in a simpler manner. The integrated application of different types of Multiple Regression Analysis (MRA) provided a cutting-edge approach to 97 replicate peak flow rates and predict flooding probabilities in sewer-catchments and opens new 98 99 lines of research and in the field of urban water planning and management. The usefulness of 100 the proposed methodology was evaluated through a case study of an urban catchment located 101 in Espoo (Finland), which provided the precipitation and flow data required to validate the 102 results at the nodes forming its sewer network.

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104 Methodology

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Since the delineation of urban catchments using Geographic Information Systems (GIS) is a widely studied topic in literature (Knebl et al. 2005; Guan et al. 2015; Jato-Espino et al. 2016a), the methodology proposed in this research only focused on reproducing the role of stormwater models using Multiple Regression Analysis (MRA). The general purpose of MRA is to predict the value of a dependent variable or predictand based on the values of a series of independent explanatory variables or predictors.

In the context of this research, Multiple Linear Regression (MLR) was used as an exploratory analysis to identify relevant parameters to the occurrence of peak flow rates. Next, the identified parameters were incorporated into a Multiple Non-Linear Regression (MNLR) framework, in order to boost the prediction accuracy of the magnitude of runoff that might be generated in urban catchments and conveyed by sewer networks as a result of severe rainfall

117 events. The magnitude of runoff was presented through the values of Maximum Lateral Inflow 118 (MLI, 1/s) and Maximum Total Inflow (MTI, 1/s) produced in urban catchments. The former 119 stands for the peak of apportionment of surface runoff from the subcatchment areas to each 120 node in the sewer network, whilst the latter also includes the contribution from preceding nodes 121 and conduits. Finally, Multiple Binary Logistic Regression (MBLR) models were created from the estimates of MLI and MTI to determine the probability of flooding in sewer networks, 122 facilitating the adoption of measures for preventing urban flood events at strategic sites and 123 maximizing their positive impact and effectiveness. Fig. 1 provides a graphical scheme of the 124 125 proposed approach based on the sequential application of these three types of MRA.

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Fig. 1. Scheme of the proposed methodology based on Multiple Regression Analysis (MRA)

130 The development of these analyses stemmed from a set of catchment parameters 131 considered for estimating MLI, whose combination with a list of predictors related to the two main elements forming sewer networks (nodes and conduits) enabled modelling both MTI and 132

- 133 flooding probability. Table 1 links these parameters to the visual objects required to represent
- 134 urban drainage systems: subcatchments, nodes and conduits.
- 135

Table 1. Parameters for estimating Flooding (%), Maximum Lateral Inflow (MLI, l/s) and Maximum Total Inflow (MTI, l/s) in urban catchments

Predictand	Sub-predictand	Visual object	ID	Predictor
Flooding (%)	MLI (l/s)	Catchment	<i>x</i> _{1.1}	Subcatchment area (ha)
			<i>x</i> _{1.2}	Degree of imperviousness in the subcatchment (%)
			<i>x</i> _{1.3}	Subcatchment width (m)
			<i>x</i> _{1.4}	Average slope in the subcatchment (%)
	MTI (l/s)	Sewer network	<i>x</i> _{2.1}	Node invert elevation (m)
			<i>x</i> _{2.2}	Preceding length of conduit (m)
			<i>x</i> _{2.3}	Cumulative preceding length of conduits (m)
			<i>x</i> _{2.4}	Subsequent length of conduit (m)
			<i>x</i> _{2.5}	Preceding diameter of conduit (m)
			<i>x</i> _{2.6}	Cumulative preceding diameter of conduits (m)
			<i>x</i> _{2.7}	Subsequent diameter of conduit (m)
			<i>x</i> _{2.8}	Preceding slope of conduit (%)
			<i>x</i> _{2.9}	Cumulative preceding slope of conduits (%)
			<i>x</i> _{2.10}	Subsequent slope of conduit (%)

These predictors were selected to result in a set of basic variables easy to acquire and/or 139 140 compute, in order to facilitate the implementation of the proposed methodology worldwide. Therefore, the length and depth of conduits (pipes) and nodes (manholes), respectively, were 141 the inputs required to parameterize the drainage network in the study area, whilst the 142 subcatchments forming it were characterized according to their area, percentage of 143 imperviousness, slope and width, which was estimated dividing their area by the average length 144 of the flow paths from the furthest drainage points. All these catchment-related variables were 145 146 determined using GIS-based editing and zonal statistics tools.

As pointed out by Yao et al. (2017), who highlighted the complexity of urban hydrology for water resources planning and management, the relationships between rainfall-runoff processes and spatial patterns is a key factor to manage flood risks in small catchments. In fact, the need for designing strategies for flood risk prevention based on proactive spatial planning has been identified as a crucial aspect to ensure the resilience of urban socio-ecological systems(Hegger et al. 2016).

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154 Identification of relevant parameters for peak flow generation

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The justification of which parameters related to the morphology of urban catchments and the geometry of sewer networks (Table 1) contributed more to explaining the values of MLI and MTI associated with different storm events was carried out using MLR. MLR consists of modelling the relationship between *n* predictors x_n and a predictand *y* through a linear expression as formulated in Eq. (1) (Aiken et al. 2003):

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$$y = b_0 + \sum_{i=1}^n b_i \cdot x_i + \varepsilon$$
(1)

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163 where b_i is the weight that indicates the importance of each predictor in the model. All 164 the information that cannot be provided by the independent variables is completed by b_0 and 165 ε , which are a constant and the residuals, respectively. A significance level of $\alpha = 0.05$ (Fisher 166 1925) was the threshold which determined whether a parameter was statistically significant for 167 estimating MLI and MTI or not.

168 Therefore, this first step had the sole purpose of justifying the selection of parameters 169 that contributed the most to reach high values of MLI and MTI from a physical point of view. 170 The use of MLR was deemed essential to ensure the hydrological and hydraulic validity of the 171 proposed approach, since the clarity of these contributions might be distorted if evaluated 172 through MNLR, due to the increased complexity involved by the inclusion of non-linear terms. 173 This course of action was adopted based on the results of similar previous studies (Jato-Espino et al. 2017), which applied MNLR to refine the prediction accuracy of relevant parametersidentifiable using MLR.

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177 Modelling of Maximum Lateral Inflow and Maximum Total Inflow

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Once the identification of catchment and sewer network parameters proving to be statistically significant for the generation of peak flow rates was accomplished, MNLR was used to combine them in the creation of non-linear equations for predicting both MLI and MTI with high accuracy. Unlike Eq. (1), the formula associated with MNLR also includes interactions and quadratic terms, as shown in Eq. (2):

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$$y = b_0 + \sum_{i=1}^{n} b_i \cdot x_i + \sum_{i=1}^{n} b_{ii} \cdot x_i \cdot x_i + \sum_{i=1}^{n} \sum_{j=1}^{n} b_{ij} \cdot x_i \cdot x_j + \varepsilon$$
(2)

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186 where b_{ii} and b_{ij} are the weights that indicate the importance of the quadratic and 187 interaction terms in the model.

Since the objective of this paper was the design of a methodology to assess flood risk in urban catchments for water resources planning and management, the predicted R^2 coefficient was chosen to measure the goodness-of-fit of MNLR, in order to validate the proposed approach for modelling new cases. This coefficient consists of removing each observation from the dataset in the MNLR, predicting the regression equation and calculating how well the model estimates the omitted observation.

194 To further ensure the validity of the MNLR models built, their residuals were analyzed 195 based upon the assumptions of normality, homoscedasticity and independence. The fulfillment 196 of these assumptions was graphically verified using the following residual plots (Osbourne and 197 Waters 2002): normal probability plot (Q-Q plot), standardized residuals vs. standardized
198 predicted values plot and standardized residuals vs. observation order plot.

199 The equations derived from these MNLR were used to estimate the values of MLI and 200 MTI in urban catchments straightforward through a series of predictors as those listed in Table 1. Such equations were further processed to incorporate rainfall intensity (1) into them, in order 201 202 to allow predicting the magnitude of flooding for different precipitation scenarios accordingly. To this end, MLR was used again to determine the constant b_0 and weights b_i , b_{ii} and b_{ij} of the 203 terms in the MNLR models as a function of I through Eq. (3). The estimates obtained using 204 205 this expression were inputs in the calculation of the flooding probability of the nodes forming urban sewer networks through MBLR. 206

(3)

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$$b_0 \vee b_i \vee b_{ii} \vee b_{ii} = b'_0 + b'_1 * b_1$$

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209 Prediction of flooding probability

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MBLR was used to integrate the knowledge generated through the application of the other types of MRA considered in the methodology (MLR and MNLR) to assess flood risk in urban catchments, MBLR is another variant of MLR characterized by the dichotomous nature of the predictand, which only has two possible outcomes. Hence, MBLR predicts the probability Pr that a certain characteristic is present in the predictand *y* from the values of the predictors x_i . The analytic expression for a MBLR model based on the logit link function is given in Eq. (4):

$$\Pr(y = 1 \mid x_i) = \frac{\exp(y')}{1 + \exp(y')} = \frac{\exp(b_0 + \sum_{i=1}^n b_i \cdot x_i)}{1 + \exp(b_0 + \sum_{i=1}^n b_i \cdot x_i)}$$
(4)

where y = 1 indicates that the characteristic is present in observation *i*. In this case, the two outcomes for the predictand were "yes" and "no", depending on whether the nodes of sewer networks were flooded or not after the occurrence of heavy rainfall events, and the predictors were the parameters included in Table 1.

In addition to the predictors and the predictand, MBLR enables incorporating another 223 term into the analysis, known as frequency, which is an indicator of the number of times that 224 the characteristic to be modelled is present. This concept was adapted to the purpose of this 225 study to express how susceptible the nodes in sewer networks were to flooding. Hence, based 226 227 on the parameters considered in the methodology so far, the frequency was defined as the ratio 228 of MTI to MLI. This value was expected to provide a measure of the sensitivity of the nodes of sewer networks to reach their full capacity, since it combines two of the main factors 229 230 favouring the occurrence of floods: accumulation and immediate contribution.

Therefore, flooding was a dichotomous dependent variable (i.e. its presence in a node is either "yes" or "no") to be estimated using a series of catchment and sewer network continuous independent variables modulated by a frequency term representing the peak flow conditions in the sewer network. Consequently, the application of Eq. (4) yielded a probability indicating how likely a certain node was to be flooded; i.e. the worse the combinations of values in the predictors and the higher ratios of MTI to MLI, the closer the probability (expressed as a decimal) of that node to be 1.

The goodness-of-fit evaluation of MBLR slightly differed from that used for MNLR, due to the particular nature of the predictand in this type of MRA. The quality of MBLR models was assessed through the adjusted deviance R^2 coefficient and the Akaike Information Criterion (AIC) (Akaike 1973), which enabled the comparison of models with different predictors. Furthermore, the Hosmer-Lemeshow test was applied to check whether the deviation between estimated and observed probabilities was unpredictable by the binomial distribution (Hosmer and Lemeshow 2000). This test was found to be more suitable than the
Deviance or Pearson tests due to the binary/response/frequency format of the data. The fact
that the predictand was a binary outcome made the verification of residuals described for
MNLR nonsensical.

The application and testing of the proposed framework enabled detecting which subcatchments and nodes required priority actions in terms of urban drainage planning and management, based on the values of peak runoff and flooding probability obtained through the subsequent application of Eqs. (2), (3) and (4).

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253 Results and discussion: a case study in Espoo, Finland

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The proposed methodology was implemented through a case study consisting of an urban 255 catchment located in Espoo, Finland (see (Sillanpää and Koivusalo 2015) for further details). 256 Fig. 2a) shows the spatial arrangement of the sewer network corresponding to this catchment, 257 which was provided by the Helsinki Region Environmental Services Authority HSY and 258 consisted of 75 nodes and 80 conduits, the 79 subcatchments forming the whole catchment 259 area, which covered 10.535 ha, and the relationship between impervious and pervious areas in 260 the catchment, which were delineated from the ortophoto of the study area (Jato-Espino et al. 261 2017). Fig. 2b) depicts the values of slope in the catchment, which were determined and 262 263 classified from the Digital Terrain Model of the study area using Geographic Information Systems (GIS) tools (Jato-Espino et al. 2016b). 264



Fig. 2. a) Sewer network, subcatchments and impervious and pervious areas in the study catchment b) Slope (%) in the study catchment
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The rainfall events used in this paper were taken from Jato-Espino et al. (2017), who modelled the study catchment in SWMM 5.1.010 (USEPA 2016) using three calibration (CAL 1, CAL 2 and CAL 3) and validation (VAL 1, VAL 2 and VAL 3) rainfall events (Table 2). These simulations reproduced the real hydrographs monitored at the outlet of the catchment with high accuracy, as demonstrated by the goodness-of-fit measures used to test them: Root-Sum Squared Error (RSSE), coefficient of determination (R^2) and Nash–Sutcliffe model efficiency coefficient (E) (Table 2).

The study catchment was re-simulated with the calibrated parameters for different return periods and Climate Change scenarios: RCP4.5 and RCP8.5 (Moss et al. 2008). Table 2 lists the values of duration, depth and intensity associated with four combinations of Climate Change scenario and return period (T) producing floods of different magnitude in the

- 281 catchment, which were determined through its lag time and the coupling of Intensity–Duration–
- 282 Frequency (IDF) curves and the Alternating Block Method, respectively.
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Table 2. Summary of the rainfall events used to test the proposed methodology. Adapted from Jato-Espino et al.
 (2017)

Event	Duration (min)	Rainfall depth (mm)	Intensity (mm/h)	RSSE	R ²	Ε
CAL 1	352	5.0	0.85	81.94	0.91	0.85
CAL 2	686	37.4	3.27	212.81	0.93	0.86
CAL 3	418	12.2	1.75	92.67	0.96	0.93
VAL 1	396	5.2	0.79	42.46	0.97	0.97
VAL 2	288	9.0	1.88	68.26	0.95	0.92
VAL 3	408	23.4	3.44	115.64	0.97	0.96
RCP4.5; T = 5 yr.	106	19.0	10.75	-	-	-
RCP8.5; T = 5 yr.	106	25.6	14.48	-	-	-
RCP4.5; T = 50 yr.	106	31.5	17.84	-	-	-
RCP8.5; T = 25 yr.	106	38.0	21.51	-	-	-

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287 Identification of relevant parameters for peak flow generation in the study

288 catchment

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Multiple Linear Regression (MLR) enabled identifying which predictors listed in Table 1 were statistically significant for the generation of peak flow rates and, by extension, determining their degree of contribution to producing high values of Maximum Lateral Inflow (MLI) and Maximum Total Inflow (MTI). MLR models were built stepwise in Minitab 17 (Minitab Inc 2016) to select only those predictors that were statistically significant to explain variations in MLI and MTI at the 95% confidence level (p-value < 0.05).

The results obtained for the modelling of MLI revealed that three parameters were statistically significant for estimating this predictand (p-value < 0.05): $x_{1.1}$ (Subcatchment area), $x_{1.2}$ (Degree of imperviousness in the subcatchment) and $x_{1.4}$ (Average slope in the subcatchment). The most influential predictor for estimating MLI was found to be $x_{1.1}$ with an average contribution of 82.52%, followed by $x_{1.2}$ and $x_{1.4}$ with 6.03% and 2.04%, respectively. 301 Although their weights were different depending on the characteristic of the rainfall events 302 used (Table 2), the contribution of the predictors considered was very similar in all cases. This 303 homogeneity of values under different rainfall events validates the results achieved, since it indicates that the impacts of $x_{1.1}$, $x_{1.2}$ and $x_{1.4}$ on MLI were very similar both when considering 304 common (CAL 1, CAL 2, CAL 3, VAL 1, VAL 2 and VAL 3) and extreme storms. 305 Furthermore, the relationships between these predictors and MLI were logical, because larger 306 subcatchments provide more opportunities to accumulate runoff and both impervious and steep 307 areas facilitate the rapid conveyance of water. Hence, areas devoid of divisions due to drainage 308 309 network deficiencies, built-up surfaces and topographically problematic sites were found to be 310 more prone to produce high lateral inflows in urban catchments. Consequently, the mitigation of excessive runoff should be approached merging both nature and artificial solutions aimed at 311 vegetating urban surfaces and also ensuring drainage support services, respectively. 312

Since the methodology was a stepped process in which the values of MTI in the nodes 313 were partially calculated from those of MLI, the latter was included in the MLR models for 314 315 estimating the former as a single predictor, in addition to those related to the sewer network 316 (Table 1). As a result, x_1 (MLI) emerged as one of the two parameters proving to be statistically significant for predicting MTI, along with $x_{2.3}$ (Cumulative preceding length of conduits). In 317 this case, $x_{2.3}$ was clearly the most important factor influencing the values of MTI in the nodes 318 of the study catchment, with an average contribution of 97.39%. Again, both predictors were 319 positively correlated to the predictand, since they contributed to increased runoff accumulation 320 321 throughout the sewer network and the catchment, respectively. In fact, x_1 was statistically 322 significant only for the extreme events. However, since they represented the situations in which 323 floods occurred for different combinations of climate scenario and return period, this parameter 324 was concluded to be relevant for the purpose of this research and was therefore not removed 325 from further analyses.

All these results enabled validating the MLR models built for identifying statistically significant parameters for estimating peak flow rates and, therefore, using the information related to their degree of contribution to create MNLR models to predict MLI and MTI with high accuracy.

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331 Modelling of Maximum Lateral Inflow and Maximum Total Inflow in the

332 study catchment

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Multiple Non-Linear Regression (MNLR) was used to fit the values of MLI and MTI obtained in all the nodes of the study catchment (Fig. 2a)) through the simulations run in SWMM (Jato-Espino et al. 2017) for the rainfall events that produced floods in the study catchment (Table 2), based on the knowledge acquired from the application of MLR to determine which catchment and sewer network parameters were more relevant for producing peak flow rates. Table 3 and Table 4 summarize the MNLR models determined for predicting both MLI and MTI.

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Table 3. Summary of the Multiple Non-Linear Regression (MNLR) models built for the estimation of Maximum
 Lateral Inflow (MLI, l/s)

Event		Equation	Pred. R^2
RCP4.5	5; T = 5 yr.	$MLI = (3.21 - 54.85 * x_{1.1} * x_{1.1} + 1.64 * x_{1.1} * x_{1.2} + 1.51 * x_{1.1} * x_{1.4})^{1/0.845}$	0.95
RCP8.5	5; $T = 5$ yr.	$MLI = (4.17 - 75.00 * x_{1.1} * x_{1.1} + 2.29 * x_{1.1} * x_{1.2} + 2.04 * x_{1.1} * x_{1.4})^{1/0.845}$	0.96
RCP4.5	5; T = 50 yr.	$MLI = (4.94 - 93.04 * x_{1.1} * x_{1.1} + 2.90 * x_{1.1} * x_{1.2} + 2.51 * x_{1.1} * x_{1.4})^{1/0.845}$	0.96
RCP8.5	5; T = 25 yr.	$MLI = (5.77 - 113.55 * x_{1.1} * x_{1.1} + 3.60 * x_{1.1} * x_{1.2} + 3.02 * x_{1.1} * x_{1.4})^{1/_{0.845}}$	0.96

	· · ·		
	Event	Equation	Pred. R^2
	RCP4.5; T = 5 yr.	$MTI = (6.03 + 9.50 * 10^{-2} * x_{2.3} + 6.39 * 10^{-2} * x_1 - 2.50 * 10^{-5} * x_{2.3} * x_{2.3})^{1/0.682}$	0.96
	RCP8.5; T = 5 yr.	$MTI = (7.70 + 1.15 * 10^{-1} * x_{2.3} + 7.11 * 10^{-2} * x_1 - 4.00 * 10^{-5} * x_{2.3} * x_{2.3})^{1/_{0.682}}$	0.95
	RCP4.5; T = 50 yr.	$MTI = (8.98 + 1.30 * 10^{-1} * x_{2.3} + 7.54 * 10^{-2} * x_1 - 5.10 * 10^{-5} * x_{2.3} * x_{2.3})^{1/_{0.682}}$	0.93
	RCP8.5; T = 25 yr.	$MTI = (10.04 + 1.42 * 10^{-1} * x_{2.3} + 7.90 * 10^{-2} * x_1 - 6.20 * 10^{-5} * x_{2.3} * x_{2.3})^{1/_{0.682}}$	0.92
1	7		

Table 4. Summary of the Multiple Non-Linear Regression (MNLR) models built for the estimation of Maximum
 Total Inflow (MTI, 1/s)

The high values of predicted R^2 reached for both models, which were always above 0.9, 348 ensured their reliability for making new estimates of MLI and MTI and validated the two-step 349 approach based on the combination of MLR and MNLR. Furthermore, their residuals met the 350 assumptions on which Multiple Regression Analysis (MRA) is based: normality, 351 352 homoscedasticity and independence. For instance, Fig. 3 provide visual verification of the fulfilment of these assumptions for the worst regression models in Table 3 and Table 4 in terms 353 of goodness-of-fit: RCP4.5; T = 5 yr. (MLI) and RCP8.5; T = 25 yr. (MTI). The approximate 354 355 straight line in the Q-Q plots ensured the normality of residuals, whilst their random and non-356 curvilinear distributions around the horizontal axis in the standardized residuals versus fits plots confirmed the homoscedasticity of the regression models. Finally, the residuals versus 357 order plots suggested that there was no serial correlation and their independence could also be 358 359 assumed.



Fig. 3. Residual analyses for the Multiple Non-Linear Regression (MNLR) models built for the estimation of a)
 Maximum Lateral Inflow (MLI, I/s) for the RCP4.5 scenario and a return period of 5 years and b) Maximum
 Total Inflow (MTI, I/s) for the RCP8.5 scenario and a return period of 25 years

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To facilitate the application of the equations for estimating MLI and MTI shown in Table 3 and Table 4 according to the characteristics of rainfall events, their constant b_0 and weights b_i , b_{ii} and b_{ij} (Eq. (2)) were fitted using MLR again with the intensity (*I*) of the Climate Change storms listed in Table 2 as predictor. Table 5 collects the equations obtained to predict these weights for both MLI and MTI.

371

372Table 5. Summary of the Multiple Linear Regression (MLR) models built for the estimation of the constant $(\boldsymbol{b_0})$ 373and weights $(\boldsymbol{b_i}, \boldsymbol{b_{ii}} \text{ and } \boldsymbol{b_{ij}})$ for the Maximum Lateral Inflow (MLI, l/s) and Maximum Total Inflow (MTI, l/s)

	Variable	Equation	Pred. R^2
	MLI	$b_{1.0} = 0.692 + 0.238 * I$	1.00
		$b_{1.1*1.1} = 3.675 - 5.434 * I$	1.00
		$b_{1.1*1.2} = -0.329 + 0.182 * I$	1.00
		$b_{1.1*1.4} = 0.008 + 0.140 * I$	1.00
\mathbf{Y}	MTI	$b_{2.0} = 2.115 + 0.380 * I$	0.97
		$b_{2.3} = 0.049 + 0.005 * I$	0.96
		$b_1 = 0.046 + 0.002 * I$	0.88
		$b_{2.3*2.3} = 1.111 * 10^{-5} - 3.471 * 10^{-5} * I$	0.98

374

375 Again, the MLR models obtained highlighted by the excellent values of predicted R^2 376 achieved and enabled accepting the assumptions of MRA, as proven in Fig. 4, which depicts the residuals plots associated with the less accurate equations in Table 5: interaction between catchment area and average slope in the subcatchment $(b_{1.1*1.4})$ and b) Maximum Lateral Inflow (MLI) (b_1)

380



Fig. 4. Residual analyses for the Multiple Linear Regression (MLR) models built for the estimation of weights
 for a) interaction between subcatchment area and average slope in the subcatchment (*b*_{1.1*1.4}) and b) Maximum
 Lateral Inflow (MLI) (*b*₁)

The merger of the expressions contained in Table 3 and Table 4 with those determined in Table 5 yielded Eqs. (5) and (6), whose application allowed calculating the values of MLI and MTI in the study catchment from the sole use of easy-to-compute GIS-based factors and the intensity of the rainfall event to be assessed.

390

381

$$MLI = [(0.692 + 0.238 * I) + (3.675 - 5.434 * I) * x_{1.1} * x_{1.1} + (-0.329 + 0.182 * I) * x_{1.1} * x_{1.2} + (0.008 + 0.140) * x_{1.1} * x_{1.3}]^{1/0.845}$$
(5)

391

$$MTI = [(2.115 + 0.380 * I) + (0.049 - 0.005 * I) * x_{2.1} + (0.046 + 0.002 * I) * x_1 + (1.111 * 10^{-5} - 3.471 * 10^{-5} * I) * x_{2.1} * x_{2.1}]^{1/0.682}$$
(6)

393 The particularization of Eqs. (5) and (6) to the Climate Change storm events shown in 394 Table 2 produced the values of MLI and MTI represented in Fig. 5 and Fig. 6, respectively. 395 Their comparison with the results obtained through simulation in SWMM resulted in values of 396 R^2 higher than 0.9 in all cases, demonstrating the accuracy of the proposed framework based 397 on the combination of MLR and MNLR to fit the lateral and total peak flow rates in the nodes of the study area due to a series of storms with different intensities and durations, which in turn 398 399 enable putting a focus on the elements in urban catchments which most contribute to producing 400 flood events and taking water-related actions accordingly.





402

403
404Fig. 5. Fit between the values of Maximum Lateral Inflow (MLI, l/s) obtained through stormwater simulations
and those determined using Multiple Non-Linear Regression (MNLR)



407 Fig. 6. Fit between the values of Maximum Total Inflow (MTI, 1/s) obtained through stormwater simulations
 408 and those determined using Multiple Non-Linear Regression (MNLR)

406

410 **Prediction of flooding probability in the study catchment**

411

The last step to accomplish the implementation of the proposed methodology to the study catchment consisted of combining the parameters listed in Table 1 with the values of MLI and MTI predicted through Eqs. (5) and (6), in order to build Multiple Binary Logistic Regression (MBLR) models for predicting the probability of flooding throughout the sewer network. The equations to estimate y' (see Eq. (4)) for the four rainfall scenarios under consideration are provided in Table 6.

418

Table 6. Summary of the Multiple Binary Logistic Regression (MBLR) models built for the estimation of
 Flooding Probability (%)

Event	Equation	Adj. Dev. R ²	AIC	H-L
RCP4.5; T = 5 yr.	$y' = -0.49 - 0.26 * x_{1.2} + 0.01 * x_{2.3} + 0.06 * x_{2.4}$	0.83	84.23	0.42
RCP8.5; T = 5 yr.	$\mathbf{y}' = -0.17 - 0.26 * x_{1.2} + 0.08 * x_{2.2} + 0.01 * x_{2.3}$	0.81	80.32	0.37
RCP4.5; T = 50 yr.	$\mathbf{y}' = -26.41 - 0.13 * x_{1.2} + 0.41 * x_{1.4} + 6.23 * x_{2.1} + 0.11 * x_{2.2} + 18.39 * x_{2.5}$	0.79	88.41	0.07
RCP8.5; T = 25 yr.	$\mathbf{y}' = -27.83 + 11.09 * x_{1.1} + 9.79 * x_{2.1} + 22.13 * x_{2.5} - 0.29 * x_{2.9}$	0.74	106.36	0.19

The quality of these models was ensured by the high and low values of adjusted 422 deviance R^2 and Akaike Information Criterion (AIC) reached, which guaranteed the suitability 423 424 of the predictors included in the equations presented in Table 6 to predict flooding probabilities. 425 Moreover, the results of the Hosmer-Lemeshow (H-L) test for the four MBLR models built yielded p-values above the significance level ($\alpha = 0.05$) in all cases, which further validated 426 their goodness-of-fit. Consequently, the inclusion of the ratio of MTI to MLI as the frequency 427 in MBLR demonstrated to be a key factor to enhance the fit between predicted and simulated 428 429 probabilities of flooding.

The values obtained for MLI and flooding probability were imported to GIS and 430 431 mapped as depicted in Fig. 7. MLI was represented according to the subcatchments based on their peak runoff rates, in order to determine their degree of contribution to the nodes to which 432 they flowed, whilst flooding probability was illustrated through the nodes forming the sewer 433 network under study. Since the Climate Change events used to test the methodology provided 434 the ranges of values for which drainage systems will have to be designed in the future to prevent 435 the occurrence of floods in the catchment, this map provided the information required to plan 436 water management strategies to take priority action in vulnerable areas. 437



439

440 **Fig. 7.** Maximum Lateral Inflow (MLI, 1/s) and Flooding Probability (%) in the subcatchments and nodes in the study area

The results represented in Fig. 7 for short return periods (RCP4.5; T = 5 yr. and RCP8.5; T = 5 yr.) indicated that oversizing the existing sewer network would not result in a relevant improvement of the drainage capacity of the catchment, since aspects like the depth and diameter of its nodes and conduits were not significantly correlated to its flooding susceptibility 447 from a statistical point of view (Table 6). On the contrary, the implementation of Sustainable 448 Drainage Systems (SuDS), also known as BMPs, Low Impact Development (LID) or Water Sensitive Urban Design (WSUD), in those areas with higher values of MLI might decrease the 449 450 degree of imperviousness of these subcatchments and therefore reduce the amount of lateral 451 inflow received by the nodes of the sewer network too. Di Matteo et al. (2017) and Meerow and Newell (2017) highlighted the importance of the spatial distribution of SuDS to improve 452 urban water-related decision-making processes, in order to maximize their impact by locating 453 them at those sites which most contribute to produce flooding, as illustrated in Fig. 7. In fact, 454 455 the results presented in Jato-Espino et al. (2016b) demonstrated that the installation of 456 Permeable Pavement Systems at the critical areas shown in Fig. 7 prevented the occurrence of floods in the study catchment when simulating the rainfall scenarios from which these 457 458 phenomena started to occur in the catchment.

Although the common return periods used to design urban drainage systems range from 459 2 to 10 years under the assumption of stationarity (Jato-Espino et al. 2016b), Climate Change 460 is expected to accelerate the water cycle in Finland, producing earlier peak flows and increased 461 discharges (Korhonen and Kuusisto 2010). Therefore, exploring the potential consequences 462 derived from storms associated with longer return periods (RCP4.5; T = 50 yr. and RCP8.5; T 463 464 = 25 yr.) must be a first concern too. According to Table 6, these scenarios would require taking integrated solutions based on extending the capacity of the existing drainage network through 465 larger diameters and invert elevations and smoother slopes, whilst implementing alternative 466 measures to complement its efficiency, such as installing SuDS and/or including new nodes to 467 468 divide existing subcatchments into smaller areas and reduce high inflow rates in some nodes. 469 The maps illustrated in Fig. 7 can be of great help for focusing on critical areas and optimize 470 the planning and management of resources to prevent floods.

472 **Conclusions**

473

This paper proposed and validated a methodology based on Multiple Regression Analysis (MRA) for assessing flood risk in urban catchments. Multiple Linear Regression (MLR) was applied to select catchment and sewer networks parameters proving to be influential in the occurrence of runoff peaks in urban areas, whilst Multiple Non-Linear Regression (MNLR) and Multiple Binary Logistic Regression (MBLR) models were built to make predictions of Maximum Lateral Inflow (MLI) and Maximum Total Inflow (MTI) in urban catchments and determine the probability of flooding across them, respectively.

The excellent values reached in the MNLR models for the predicted coefficient of 481 482 determination proved that the combination of catchment and sewer network parameters, especially subcatchment area and cumulative length of preceding conduits, can provide 483 484 accurate estimates of the maximum peak flow rates in subcatchments and nodes. The subsequent use of MBLR provided high-accuracy prediction models to determine the flooding 485 probability associated with the nodes of sewer-catchments under different extreme rainfall 486 scenarios produced by Climate Change. The results proved that the implementation of 487 Sustainable Drainage Systems (SuDS) might be enough to mitigate floods for the return periods 488 489 commonly used for urban designs, whilst integrated approaches combining both conventional 490 and alternative water management measures would be required to deal with more extreme scenarios. The fact that these outcomes were based on easy to acquire and/or produce 491 492 parameters and their relationships were solid in both physical and statistical terms enabled the 493 extrapolation and generalization of the proposed approach to other case studies, since the 494 interpretation and application of MRA is simple and compatible with Geographic Information 495 Systems (GIS).

496 This methodology is presented as an accessible framework to support the adoption of 497 measures by administrative entities for facilitating their drainage management planning actions 498 and maximizing their impact through their implementation at strategic sites in terms of flood 499 susceptibility. Although the reliability of the results to which it led is not compromised by the 500 location of the study area, further research should consider the application of this methodology to other urban catchments with larger areas, different climate conditions and more complex 501 drainage systems, in order to regionalize the development of prediction models according to 502 503 the degree of similarity of distinct zones worldwide. The other main future line of action to 504 continue this research should consist of exploring the automation of the proposed methodology 505 through easy-to-use interfaces and/or support tools, so that potential decision-makers and water 506 resources planners without expertise in the statistical techniques considered might apply them 507 by merely providing a series of basic weather and physical inputs.

508

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510

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